1	Re	source partitioning based on body size contributes to the species diversity of wood-boring
2	bee	etles and arboreal nesting ants
3		
4	Tos	shiyuki Satoh ¹ , Tomohiro Yoshida ^{2,*} , Satoshi Koyama, Akira Yamagami ³ , Mamoru Takata ⁴ ,
5	Hay	yato Doi ⁴ , Takuma Kurachi ⁵ , Shinya Hayashi ⁴ , Takuya Hirobe ⁴ , Yasuo Hata ⁴
6		
7	¹ Ins	stitute of Agriculture, Tokyo University of Agriculture and Technology, Fuchu, Japan
8	² Fa	culty of Agriculture, Field Science Center, Tokyo University of Agriculture and Technology,
9	Fuc	chu, Japan
10	³ Re	esearch Institute of Civilization, Tokai University, Hiratsuka, Japan
11	⁴ Fa	culty of Agriculture, Tokyo University of Agriculture and Technology, Fuchu, Japan
12	⁵ Ur	nited Graduate School of Veterinary Science, Gifu University, Gifu, Japan
13		
14	*Co	orresponding author
15		E-mail address: yoshitom@cc.tuat.ac.jp
16		
17	Ab	stract
18	1.	The need to understand the mechanisms enhancing species coexistence and diversity is one of
19		the central issues of community ecology. Theoretical work suggests that resource partitioning
20		based on body size contributes to the coexistence of competing species, resulting in increased
21		species diversity. Nevertheless, few empirical studies have investigated this. The species
22		composition of wood-dwelling arthropods was surveyed to examine resource partitioning based
23		on body size in the dead wood of the ubame oak (Quercus phillyraeoides A. Grey) trees.
24	2.	Greater resource sizes (larger dead wood diameter) retained a lower overall density of
25		wood-dwelling beetles per resource volume, greater total beetle biomass per resource volume
26		and larger bodied individuals of the dominant species (Cerambycidae; Anaglyptus nipponensis
27		Bates). A gradient in the species composition of arboreal ant assemblages along resource size
28		and a positive correlation between resource size and ant body size was observed. These results
29		suggest that the incidence of wood-boring beetles was limited by the size of the available
30		resource, resulting in resource partitioning by the ants according to the sizes of galleries made
31		by the beetles.
32	3.	Our findings indicate that resource size leads both directly and indirectly to body
33		size-dependent resource partitioning in beetle and ant assemblages, and contributes to the
34		maintenance of species coexistence in the assemblages present in dead wood.
35	Ke	ywords
36	har	vester ant, intraspecific competition, mating partner
37		

l

38 Introduction

39 One of the central goals of community ecology is to understand the mechanisms that enhance 40species coexistence and support species diversity (Agrawal et al., 2007). The traditional ecological 41theory predicts that niche differentiation allows for the coexistence of potentially competitive species, 42resulting in increased species diversity (reviewed in Chesson, 2000; Chase & Leibold, 2003). Body 43size variation is one of the factors that allow competing species to avoid direct overlaps in resource 44use (Schoener, 1974; Werner & Gilliam, 1984). Differences in resource utilisation, that is, resource partitioning, between sympatric species depending on their body size promote the coexistence of 45different species living in the same habitat (MacArthur, 1972; Schoener, 1974; Bowers & Brown, 46 1982). 4748 Dead wood provides a favourable habitat to various arthropods (Grove, 2002; Stokland et al.,

492012), and is a good system in which to study the mechanisms of species coexistence and diversity. 50The size (i.e. diameter) of dead wood is a determinant of the community structure of wood-dwelling 51arthropods. Past studies have shown that an increase in the size of dead wood leads to an increase in 52the average body size of the beetles (Hanks et al., 2005) and their abundance (Banno & Yamagami, 531991), and to a change in the species compositions of their assemblages (Jonsell et al., 2007; Foit, 542010). The increase in resource size provides a separate niche for primary colonisers such as beetles, 55which are the first to utilise the inside and surface of dead wood. The abandoned galleries and 56tunnels made by the primary colonisers subsequently have a positive effect on the species richness 57and abundances of secondary colonisers and their predators (Buse et al., 2008; Calderón-Cortés et al., 582011). The primary colonisers, which create habitats for other organisms, are called ecosystem 59engineers (Jones et al., 1994, 1997). This is recognised to be an important group, which constructs 60 'indirect interaction webs' (Ohgushi, 2005).

Wood-nesting ants are secondary colonisers, which utilise abandoned galleries and tunnels made by wood-boring beetles within dead wood (Tschinkel, 2002; Powell, 2008; Powell *et al.*, 2011). Among wood-dwelling ants, the occurrence of *Camponotus* was greater in dead wood with a larger diameter size than in that with the smallest (Torgersen & Bull, 1995). This means that the traits of wood-boring beetle assemblages (body size, abundance and species composition) affect the composition of wood-nesting ant assemblages through the configuration of their abandoned galleries

- and tunnels. Nevertheless, few studies have been conducted from this point of view.
- In this study, we aim to elucidate the importance of body size-based resource partitioning by wood-dwelling arthropods as the mechanism of species coexistence and diversity. We have the following two hypotheses: (i) resource size (wood diameter) directly determines the body size of wood-boring beetles (primary colonisers); (ii) resource size indirectly promotes resource partitioning by wood-nesting ants (secondary colonisers) of different body sizes through the sizing of the abandoned beetle galleries and tunnels, resulting in species coexistence within each assemblage

- 74 present in dead wood.
- 75

76 Materials and Methods

77 Study area

78 The research was conducted in a coastal forest ($34^{\circ}39'47''$ N, $138^{\circ}56'15''$ E; about 200 \times 100 m) at 79 Shimoda, Shizuoka Pref., at the southern end of the Izu Peninsula in Japan. This area is in a warm 80 temperate region under the influence of the Asia monsoon climate. The mean annual rainfall and the 81 annual mean air temperature were 1780 mm and 16.6 °C, respectively, at the nearest weather station 82 [1981–2010, Japan Meteorological Agency (JMA), 2015]. The forest was dominated by the evergreen broad-leaved ubame oak (Quercus phillyraeoides A. Grey), which is distributed along the 83 84 Pacific coast of south to central Japan, Korea and China. The ubame oak reaches a height of 5-7 m 85 and a diameter of 20 cm, and is characterised by the growth of several trunks stemming out from the 86 root. Dead trunks are abundant in the forest, likely due to damage caused by seawater and strong 87 salty winds blown in from the sea (Ito & Gyokusen, 1996), especially during the typhoon season. In 88 addition, light competition within the clump of multiple trunks (self-thinning), generally observed 89 among sprouting species, might affect mortality rates (O'Hara & Berrill, 2010). High levels of 90 exposure to salt deposition potentially have a negative effect on ground-dwelling insects (Kaspari et 91 al., 2008). Nevertheless, this would not seriously affect resource utilisation by the wood-dwelling 92arthropods studied here, because these groups mainly utilise tree trunks or the space under bark. In 93addition, because seawater deposition by a typhoon is a short-term event, the seawater would be 94 flushed from standing tree trunks by heavy rainfall.

95

96 Collection of dead wood

97 The collection of dead trunks was performed in mid-March, 1986, when the insects inhabiting the

98 trees had not yet emerged. Twelve dead trunks c. 2–3 m in height were cut down at their base,

99 stripped of their side branches and transported to the laboratory. The dead trunks were harvested

- 100 from different individuals, and were either main or sub-trunks, depending on the trees. In the
- 101 laboratory, the dead trunks were each cut into sections of length of 40 cm, and their diameter was
- 102 measured at the centre with vernier callipers. The pieces of dead wood were sorted into eight
- 103 thickness categories according to diameter (Table 1: small to large; A-H). All bottom pieces were

104 classified as the thickest group (H), and used to breed arthropods in a bucket (see below). For other

- 105 wood thickness groups (A–G), the pieces of dead wood in each thickness category were selected to
- 106 prevent sampling bias, because of the limitation of the buckets' volume. The average diameters
- 107 ranged from 5.4 mm (A) to 78.4 mm (H). The total volume of dead wood was c. 0.12 m^3 , and the
- 108 trunk section diameters ranged from 1 to 95 mm (overall average: 18.2 mm).

110 Rearing of arthropods in dead wood

111 The pieces of dead wood were placed into each 'breeding bucket' and kept at room temperature. The 112breeding buckets were each comprised of a stainless steel basket of 5-mm mesh placed in a 70-L 113 bucket (Fig. S1). A packing sponge was inserted between the glass lid and the bucket to prevent any 114 insects from escaping, while still allowing for air permeability. The moisture was supplied 115occasionally within the buckets by spraying water mist to prevent dry our completely, especially for 116 smaller dead wood groups. We checked the buckets every week. When we found emergent beetles or 117dwelling ants, we identified them to the species level. Furthermore, once every 2 weeks we sifted the 118 woody debris in the bucket with a 0.5-mm mesh sieve and identified any arthropods remaining on 119 the sieve. The study continued for 6 months, until September 1986, when we surveyed all beetles 120and ants remaining inside the dead wood. Insect carcasses found in the wood were also identified to 121the order or family level. 122123Measurement of body weight and size 124Emergent beetles were completely dried in the desiccator for more than 1 month. The body weights

125 of the beetles were measured individually using a balance (0.01 mg scale), and from these

126 measurements the average body weight and first to third quartile values were calculated for each

127 species. The head widths of the worker ants (minor workers for dimorphic species) were measured

128 with an ocular micrometre (0.01 mm scale), from which mean and SD were calculated for each

129 species. The samples were randomly selected from the collected specimens.

130

131 Data analysis

132 The species number and diversity of wood-dwelling arthropods were not compared among the wood

thickness groups, because the total weight and volume of dead wood were different among them

134 (Table 1). Abundance data were carefully interpreted when being compared among the thickness

135 groups. For all arthropods collected, the Spearman's rank correlation between the diameter of dead

136 wood and (i) total individual density and (ii) total biomass per dead wood volume was calculated.

137 The differences in the mean body weight of individuals among the dead wood groups were analysed

138 using analysis of variance (anova). Tukey's honest significant difference test was used for multiple

139 comparisons. Similarities in ant species composition between groups were calculated using Odum's

140 similarity index (Odum, 1950), and cluster analysis was performed using group average methods.

141

142 **Results**

143 Arthropod fauna

144 In total, three classes (Arachnida, Diplopoda and Insecta) and 12 orders of arthropods were reared

145 from dead ubame oak wood (Table 2). In Insecta, species of the order Hymenoptera, including ants

146 and wasps, were the most abundant, followed by species of Coleoptera, Lepidoptera and Psocoptera.

147 In addition to Insecta, Polyxenida (class: Diplopoda) was also abundant. Species of Hymenoptera,

148 Coleoptera, Psocoptera and Polyxenida appeared in all thickness groups, whereas species of

149 Psocoptera and Polyxenida appeared mainly in the small to medium thickness groups (B–D).

150 Lepidopterans appeared in the medium to large thickness groups (D–H), especially in the largest

- 151 group (H).
- 152

153 Wood-boring beetles

154 In total, 16 families, 28 genera and 29 species of Coleoptera were found in the dead wood (Table 3).

155 The density of individual beetles was higher in the small to medium thickness groups (A–E) than in

156 the trunks with larger diameters, although lager thickness groups included a larger volume and dry

- 157 weight of dead wood. The total number of individual beetles per wood volume was negatively
- 158 correlated with the diameter of the dead wood (Fig. 1a; $r_s = -0.881$, d.f. = 6, P < 0.01), whereas the
- total biomass per volume of dead wood was positively correlated with the diameter of the dead wood
- 160 (Fig. 1b; $r_s = 0.690$, d.f. = 6, P < 0.05). The dominant species were *Gastrallus affinis* Sakai

161 (Anobiidae) and Anaglyptus nipponensis Bates (Cerambycidae), of which 322 and 138 were

162 observed in total respectively (Table S1). Gastrallus affinis was abundant in the small to medium

163 thickness groups (B-E), whereas A. nipponensis was abundant in the medium to large dead wood

164 groups (B–H). Anaglyptus nipponensis individuals from the smaller dead wood (B) were

165 significantly smaller than those that emerged from the larger dead wood (C-E)

166 (Fig. 2; anova, F = 3.64, P < 0.005).

167

168 Arboreal ants

- 169 In total, three subfamilies, four genera and seven species of arboreal ants were found in the dead
- 170 wood (Table 4; Table S2). Although Tapinoma sp. (Dolichoderinae), Camponotus
- 171 *nipponicus* Wheeler (Formicinae) and *Camponotus nawai* Ito were collected from almost all dead
- 172 wood groups (Fig. 3; Table S2), they dominated different size groups. *Tapinoma* sp. dominated the
- 173 first and second smallest dead wood groups (A and B), whereas C. nipponicus dominated the third
- 174 smallest group (C), and C. nawai dominated the larger groups (D-G). Camponotus
- 175 devestivus Wheeler emerged and dominated only the largest group (H). Cluster analysis of the

176 similarities in ant community species composition revealed three major clusters corresponding to the

- 177 diameter of dead wood (Fig. 4). The first cluster, corresponding to small dead wood groups (A-C),
- 178 was characterised by the dominance of *Tapinoma* sp. and *C. nipponicus*. The second cluster,
- 179 corresponding to the medium to large groups (D–G), was characterised by the dominance
- 180 of C. nawai, and the third cluster, corresponding to group H, was characterised by the dominance

181 of C. devestivus.

 $\mathbf{5}$

182 Correspondence of body size between wood-boring beetles and arboreal ants

- 183 The body sizes of wood-boring beetles and arboreal ants increased with the diameter of the dead
- 184 wood they inhabited (Fig. 5). The smallest beetle, *Xylosandrus* sp. (Scolytidae), and the second
- 185 smallest, G. affinis, emerged mainly from trunks in the small dead wood groups and the smallest
- 186 ants, *Tapinoma* sp., and relatively small *C. nipponicus* dominated the arthropod communities in
- 187 small dead wood. The mid-sized beetle A. nipponensis emerged mainly from medium to large dead
- 188 wood, and the mid-sized ant *C. nawai* dominated the ants in those thickness groups. The largest
- 189 beetle, Mesosa longipennis Bates (Cerambycidae), emerged mainly from trunks in the largest
- 190 thickness groups, and the largest ant, *C. devestivus*, also dominated in those thickness groups.
- 191

192 **Discussion**

We showed that larger resource sizes (thicker dead wood) supported a lower total number of individual wood-dwelling beetles per resource volume, and more total beetle biomass per resource volume. This means that the biomass of wood-dwelling beetles per individual increased with increasing resource size. Smaller resource sizes supported smaller beetle species, and the smaller individuals of the dominant species (*A. nipponensis*). These results support our first hypothesis, and agree with the results of previous studies that have reported a positive correlation between dead wood size and the body sizes of wood-boring beetles (Hanks *et al.*, 2005).

200 Some parameters of dead wood, which correlate with wood diameter, affect the growth and 201survival of wood-dwelling arthropods. The differences in wood diameter provide different 202heterogeneous habitats and allow for different colonisation processes (Grove, 2002; 203Heilmann-Clausen & Christensen, 2004). The proportions of wood components, such as bark, 204cambium, phloem and sapwood, to volume are partially dictated by the diameter of the dead wood 205(Grove & Forster, 2011). Past studies showed positive correlations between wood diameter and 206 phloem thickness, and between phloem thickness and the gallery construction rate of bark beetles 207(Scolytidae) (Amman, 1977; Haack et al., 1984). Smaller diameter dead wood provides a more 208variable microclimate owing to its larger amount of surface area per volume and to smaller quantities 209of substrates that are well-buffered from the external environment (Schiegg, 2001). Zhang et al. 210(1993) demonstrated a positive correlation between trunk diameter and bark thickness, and between 211bark thickness and the emergence density of a long-horned beetle species. Nevertheless, the smaller 212diameter dead wood is made up of many more individual wood pieces per volume than the larger 213diameter wood, resulting in higher microhabitat diversity per volume (Schiegg, 2001). This causes 214the colonisation processes by wood-dwelling organisms to be more stochastic (Heilmann-Clausen &

- 215 Christensen, 2004). Therefore, smaller diameter wood provides higher colonisation opportunity for
- 216 wood-dwelling organisms, but lower survival condition due to drought. In our study, higher
- 217 emergent beetle density in smaller diameter wood would be attributed to moderate moisture

218condition by spraying water within rearing containers.

219We identified a gradient in the species composition of arboreal ants along their resource size, 220and a positive correlation between the diameter of the dead wood and the body size of the ants. 221Positive correlations between resource size and the sizes of wood-dwelling beetles and ants support 222our second hypothesis, which suggested that the different sizes of wood-boring beetles (primary 223colonisers), as dictated by their resource size, leads to the production of different sizes of galleries 224and tunnels, ultimately leading to resource partitioning by arboreal ants (secondary colonisers). Pratt 225and Pierce (2001) examined three geometric parameters of hollow acorn nests associated with nest 226site selection by cavity-dwelling ant species (Leptothorax curvispinosus): entrance hole area, cavity 227 volume and cavity shape. They clarified that this ant species preferred the cavities that had small 228entrance holes, larger volumes, and compact, high-ceilinged shapes. The wood-dwelling ants in our 229study were also influenced by these gallery and tunnel parameters.

230The entrance hole area of ant nests affects the ants' ability to defend their nests against 231predators and other ants (Powell, 2009). In some cavity-dwelling ant species, whose nest entrances

232are blocked by the heads of one or more major workers (Hölldobler & Wilson, 1990; Powell, 2008),

233the optimal entrance area is equal to the head area of the major workers of that species

234(Powell, 2009). Because the optimal entrance area differs among cavity-dwelling ant species,

235cavity-entrance diversity promotes the niche differentiation of the species (Powell et al., 2011).

236Cavity-dwelling ant species often reduce the size of the entrance area by adding rims of mud and

237debris (Herbers & Banschbach, 1995; Pratt & Pierce, 2001). Among the

238arboreal Camponotus species, the subgenera Colobopsis, Myrmamblys and Paramyrmamblys are

239known to exhibit behavioural and morphological adaptations relating to the blocking of nest

240entrances (Klimes & McArthur, 2014). The Camponotus nipponicus, C. nawai,

241and C. kiusiuensis collected in our study belong to these three subgenera, and C. nipponicus were

242observed to exhibit the behaviour of plugging nest entrances holes with their heads

243(Hasegawa, 1993). C. nipponicus and C. nawai exhibit both entrance-blocking and

244entrance-reduction behaviours (T. Satoh, pers. obs.). They could utilise a relatively wide range of

245dead wood sizes because of these behaviours. Therefore, the variability in nest entrance areas with

246the body sizes of wood-boring beetles account for the presence different ant species and their ability

247to coexist in dead wood.

248

The volume and shape of the beetles' galleries and tunnels also affect the colonisation of dead 249wood by arboreal ants. In some ant-plants (myrmecophytes), a positive correlation between domatia

250volume and ant body size causes each domatium to be occupied by a different ant species, resulting

251in the coexistence of multiple ant species on individual plants (Fonseca, 1999; Campbell et al., 2013).

- 252The volume and shape of beetles' galleries and tunnels vary with tree species, beetle species and
- 253their colonisation density (e.g. Connor & Wilkinson, 1983; Zhong & Schowalter, 1989;

Schowalter *et al.*, 1998). Cerambycid larvae have been known to excavate bark beetle mines to feed on bark beetle larvae (Coulson *et al.*, 1980; Dodds *et al.*, 2001; Schowalter, 2006). Such interactions might increase the complexity of the gallery and tunnel structures, providing a wider range of suitable habitats to secondary colonisers. Therefore, this variability in wood volume and shape also plays an important role in the determination of the species composition of wood-dwelling ant communities in dead wood.

260In conclusion, we have demonstrated that variation in resource size leads both directly and 261indirectly to size-dependent resource partitioning by wood-dwelling beetles and ants, and thus 262contributes to the maintenance of species coexistence in insect communities within dead wood. We 263can consider this beetle-ant relationship with be a 'nest web'. In cavity-nesting bird communities, 264several authors proposed and used this concept, which consists of primary cavity nesters (e.g. 265woodpeckers) and secondary cavity-nesting species that depend entirely on the production of 266 cavities by primary cavity nesters (Martin & Eadie, 1999; Cockle et al., 2012). Shelter building by 267wood-boring beetles might have additional linkages around dead wood. Nakamura and Ohgushi 268(2003) described an 'indirect interaction web' (Ohgushi, 2005) in which leaf shelters made by leaf 269rolling caterpillars increased the abundance of aphids on willow shoots, which in turn increased the 270abundance of aphid honeydew-harvesting ants, indirectly reducing the larval survival of the leaf 271beetles due to ant predation pressure. Similarly, arboreal ants act as important predators in arboreal 272food webs (Floren et al., 2002), where the species composition and abundance of arboreal arthropod 273communities depend on the presence or absence of ant nesting and ant community composition in 274dead wood. Further empirical research relating to the above concepts is needed in order to elucidate 275the mechanisms behind species coexistence in dead wood.

276

277 Acknowledgements

278 We greatly appreciate the Shimoda Marine Research Center, University of Tsukuba, for its

- 279 hospitality and many kinds of support. This manuscript was greatly improved by Dr Y. Basset and
- 280 the two anonymous reviewers. This work was supported by funding from the Japan Society for the
- 281 Promotion of Science KAKENHI program (grant numbers 24780144 and 25450483).
- 282

283 Competing interest

- We have no competing interests.
- 285

286 References

- Andersen Agrawal, A.A., Ackerly, D.D., Adler, F.R., Arnold, A.E., Cáceres, C., Doak, D.F., Post,
 E., Hudson, P.J., Maron, J., Mooney, K.A., Power, M., Schenske, D., Stachowicz, J., Strauss,
- 289 S., Turner, M.G. & Werner, E. (2007) Filling key gaps in population and community
 - 8

290	ecology. Frontiers in Ecology and the Environment, 5, 145–152.
291	Amman, G.D. (1977) Role of the mountain pine beetle in Lodgepole pine ecosystems: impact on
292	succession. The Role of Arthropods in Forest Ecosystem (ed. by W.J. Matton), pp. 3-18.
293	Springer-Verlag, New York City, New York.
294	Banno, H. & Yamagami, A. (1991) Life cycle and larval survival rate of the redspotted longicorn
295	beetle, Eupromus ruber (Dalman) (Coleoptera: Cerambycidae). Applied Entomology and
296	<i>Zoology</i> , 26 , 195–204.
297	Bowers, M.A. & Brown, J.H. (1982) Body size and coexistence in desert rodents: chance or
298	community structure? <i>Ecology</i> , 63 , 391–400.
299	Buse, J., Ranius, T. & Assmann, T. (2008) An endangered longhorn beetle associated with old oaks
300	and its possible role as an ecosystem engineer. Conservation Biology, 22, 329-337.
301	Calderón-Cortés, N., Quesada, M. & Escalera-Vazquez, L.H. (2011) Insects as stem engineers:
302	interactions mediated by the twig-girdler Oncideres albomarginata chamela enhance
303	arthropod diversity. PLoS ONE, 6, e19083.
304	Campbell, H., Fellowes, M.D.E. & Cook, J.M. (2013) Arboreal thorn-dwelling ants coexisting on
305	the savannah ant-plant, Vachellia erioloba, use domatia morphology to select nest sites. Insect
306	<i>Sociaux</i> , 60 , 373–382.
307	Chase, J.M. & Leibold, M.A. (2003) Ecological Niches: Linking Classical and Contemporary
308	Approaches. University of Chicago Press, Chicago, Illinois.
309	Chesson, P. (2000) Mechanisms of maintenance of species diversity. Annual Review of Ecology and
310	<i>Systematics</i> , 31 , 343–366.
311	Cockle, K.L., Martin, K. & Robledo, G. (2012) Linking fungi, trees, and hole-using birds in a
312	Neotropical tree-cavity network: pathway of cavity production and implications for
313	conservation. Forest Ecology and Management, 264, 210-219.
314	Connor, M.D. & Wilkinson, R.C. (1983) Ips Bark Beetles in the South. Forest Insect & Disease
315	Leaflet 129. U.S. Department of Agriculture, Forest Service, Washington, District of
316	Columbia.
317	Coulson, R.N., Pope, D.N., Gagne, J.A., Fargo, W.S., Pulley, P.E., Edson, L.J. & Wagner,
318	T.L. (1980) Impact of foraging by Monochamus titillator (Col.: Cerambycidae) on within-tree
319	populations of Dendroctonus frontalis (Col.: Scolytidae). Entomophaga, 25, 155-170.
320	Dodds, K.J., Graber, C. & Stephen, F.M. (2001) Facultative intraguild predation by larval
321	Cerambycidae (Coleoptera) on bark beetle larvae (Coleoptera: Scolytidae). Environmental
322	Entomology, 30 , 17–22.
323	Floren, A., Biun, A. & Linsenmair, K.E. (2002) Arboreal ants as key predators in tropical lowland
324	rainforest trees. Oecologia, 131, 137–144.
325	Foit, J. (2010) Distribution of early-arriving saproxylic beetles on standing dead Scots pine

326	trees. Agricultural and Forest Entomology, 12, 133–141.
327	Fonseca, C.R. (1999) Amazonian ant-plant interactions and the nesting space limitation
328	hypothesis. Journal of Tropical Ecology, 15, 807-825.
329	Grove, S.J. (2002) Saproxylic insect ecology and the sustainable management of forest. Annual
330	<i>Review of Ecology and Systematics</i> , 33 , 1–23.
331	Grove, S.J. & Forster, L. (2011) A decade of change in the saproxylic beetle fauna of eucalypt logs
332	in the Warra long-term log-decay experiment, Tasmania. 2. Log-size effects, succession, and
333	the functional significance of rare species. Biodiversity and Conservation, 20, 2167-2188.
334	Haack, R.A., Wilkinson, R.C., Foltz, J.L. & Corneil, J.A. (1984) Gallery construction and
335	oviposition by Ips calligraphus (Coleoptera: Scolytidae) in relation to slash pine phloem
336	thickness and temperature. The Canadian Entomologist, 116, 625-632.
337	Hanks, L.M., Paine, T.D. & Millar, J.G. (2005) Influence of the larval environment on performance
338	and adult body size of the wood-boring beetle Phoracantha semipunctata. Entomologia
339	Experimentalis et Applicata, 114, 25–34.
340	Hasegawa, E. (1993) Nest defense and early production of the major workers in the dimorphic
341	ant Colobopsis nipponicus (Wheeler) (Hymenoptera: Formicidae). Behavioral Ecology and
342	<i>Sociobiology</i> , 33 , 73–77.
343	Heilmann-Clausen, J. & Christensen, M. (2004) Does size matter? On the importance of various
344	dead wood fractions for fungal diversity in Danish beech forests. Forest Ecology and
345	Management, 201 , 105–117.
346	Herbers, J.M. & Banschbach, V. (1995) Size-dependent nest site choice by cavity-dwelling
347	ants. <i>Psyche</i> , 102 , 13–17.
348	Hölldobler, B. & Wilson, E.O. (1990) The ants. Harvard University Press, Cambridge,
349	Massachusetts.
350	Ito, S. & Gyokusen, K. (1996) Analysis of the multi-stem clump structure of Litsea japonica Juss.
351	growing in a coastal dwarf forest. Ecological Research, 11, 17–22.
352	
353	Japan Meteorological Agency (JMA). (2015) Tables of climatological normals (1981–2010).
354	< <u>http://www.jma.go.jp/jma/indexe.html</u> > 5th March 2015.
355	Jones, C.J., Lawton, J.H. & Shachak, M. (1994) Organisms as ecosystem engineers. Oikos, 69, 373-
356	386.
357	Jones, C.J., Lawton, J.H. & Shachak, M. (1997) Positive and negative effects of organisms as
358	physical ecosystem engineers. <i>Ecology</i> , 78 , 1946–1957.
359	Jonsell, M., Hansson, J. & Wedmo, L. (2007) Diversity of saproxylic beetle species in logging
360	residues in Sweden - comparisons between tree species and diameters. Biological
361	<i>Conservation</i> , 138 , 89–99.

I

- Kaspari, M., Yanoviak, S.P. & Dudley, R. (2008) On the biogeography of salt limitation: a study of
 ant communities. *Proceedings of the National Academy of Sciences of the United States of America*, 105, 17848–17851.
- Klimes, P. & McArthur, A. (2014) Diversity and ecology of arboricolous ant communities
 of *Camponotus* (Hymenoptera: Formicidae) in a New Guinea rainforest with descriptions of
 four new species. *Myrmecological News*, 20, 141–158.
- 368 MacArthur, R.H. (1972) *Geographical Ecology: Patterns in the Distribution of Species*. Harper &
 369 Row, New York City, New York.
- Martin, K. & Eadie, J.M. (1999) Nest webs: a community-wide approach to the management and
 conservation of cavity-nesting forest birds. *Forest Ecology and Management*, 115, 243–257.
- Nakamura, M. & Ohgushi, T. (2003) Positive and negative effects of leaf shelters on herbivorous
 insects: linking multiple herbivore species on a willow. *Oecologia*, 136, 445–449.
- Odum, E.P. (1950) Bird populations of the highlands (North Carolina) plateau in relation to plant
 succession and avian invasion. *Ecology*, **31**, 587–605.
- O'Hara, K.L. & Berrill, J.-P. (2010) Dynamics of coast redwood sprout clump development in
 variable light environments. *Journal of Forest Research*, 15, 131–139.
- Ohgushi, T. (2005) Indirect interaction webs: herbivore-induced effects through trait change in
 plants. Annual Review of Ecology, Evolution, and Systematics, 36, 81–105.
- Powell, S. (2008) Ecological specialization and the evolution of a specialized caste
 in *Cephalotes* ants. *Functional Ecology*, 22, 902–911.
- 382 Powell, S. (2009) How ecology shapes caste evolution: linking resource use, morphology,
- 383 performance and fitness in a superorganism. *Journal of Evolutionary Biology*, **22**, 1004–1013.
- 384 Powell, S., Costa, A.N., Lopes, C.T. & Vasconcelos, H.L. (2011) Canopy connectivity and the
- availability of diverse nesting resources affect species coexistence in arboreal ants. *Journal of Animal Ecology*, 80, 352–360.
- Pratt, S.C. & Pierce, N.E. (2001) The cavity-dwelling ant *Leptothorax curvispinosus* use nest
 geometry to discriminate between potential homes. *Animal Behaviour*, 62, 281–287.
- Schiegg, K. (2001) Saproxylic insect diversity of beech: limbs are richer than trunks. *Forest Ecology and Management*, 149, 295–304.
- 391 Schoener, T.W. (1974) Resource partitioning in ecological communities. *Science*, **185**, 27–39.
- Schowalter, T.D. (2006) *Insect Ecology, an Ecosystem Approach*, 2nd edn. Academic Press, San
 Diego, California.
- Schowalter, T.D., Zhang, Y.L. & Sabin, T.E. (1998) Decomposition and nutrient dynamics of
 oak *Quercus* spp. logs after five years of decomposition. *Ecography*, 21, 3–10.
- Stokland, J.N., Siitonen, J. & Jonsson, B.G. (2012) *Biodiversity in Dead Wood*. Cambridge
 University Press, Cambridge, UK.

398	Torgersen, T.R. & Bull, E.L. (1995) Down logs as habitat for forest-dwelling ants – the primary prey
399	of pileated woodpeckers in northeastern Oregon. Northwest Science, 69, 294-303.

- 400 Tschinkel, W.R. (2002) The natural history of the arboreal ant, *Crematogaster ashmeadi. Journal of* 401 *Insect Science*, 2, 12.
- Werner, E.E. & Gilliam, J.F. (1984) The ontogenetic niche and species interactions in size-structured
 populations. *Annual Review of Ecology and Systematics*, 15, 393–425.
- Zhang, Q.-H., Byers, J.A. & Zhang, X.-D. (1993) Influence of bark thickness, trunk diameter and
 height on reproduction of the longhorned beetle, *Monochamus sutor* (Col., Cerambycidae) in
 burned larch and pine. *Journal of Applied Entomology*, 115, 145–154.
- Zhong, H. & Schowalter, T.D. (1989) Conifer bole utilization by wood-boring beetles in western
 Oregon. *Canadian Journal of Forest Research*, 19, 943–947.





412 biomass (dry weight) of beetles respectively.



415 Figure 2. Body weight (open circle) and the number of individuals (histogram) of Anaglyptus

416 *niponensis* according to the average diameter of each thickness group of dead wood (B-H; absent in

417 group A, see Table S1). Values of the body weight represent means \pm SD. Different lower case letters

418 indicate significant difference at P < 0.05.

419



421 Figure 3. Species composition of arboreal ants in each thickness group of dead wood of ubame oak

422 (Quercus phillyraeoides).

423



425 Figure 4. Similarity of ant species composition between thickness groups of dead wood of ubame

426 oak (Quercus phillyraeoides). Similarity between groups was calculated by Odum's per cent

- 427 similarity index, and group average method was employed.
- 428



430 Figure 5. (a) Body size of wood-boring beetles (average body weight) and (b) ants (average head

431 width of workers) in relation to wood thickness: mean (open circle), first to third quartile (solid line)

432 and range (dashed line) were shown. Beetles (mean \pm SD): Mesosa longipennis (MI:

433 94.53 ± 25.32 mg, *n* = 7), *Chlorophorus muscosus* (Cm: 21.93 ± 7.31 mg, *n* = 6), *Anaglyptus*

434 *niponensis* (An: 8.94 ± 3.18 mg, n = 20), *Rhaphuma diminuta* (Rd: 2.92 mg, n = 2), *Holcobius*

435 *japonicas* (Hj: 2.51 ± 0.73 mg, n = 20), *Ptinomorphus exilis* (Pe:

436 0.89 ± 0.26 mg, n = 20), Oligomerus japonicas (Oj: 0.76 ± 0.26 mg, n = 18), Penthelispa vilis (Pv:

437 0.68 ± 0.18 mg, n = 20), *Gastrallus affinis* (Ga: 0.42 ± 0.12 mg, n = 20) and *Xylosandrus* sp. (X:

438 $0.19 \pm 0.03 \text{ mg}, n = 19$). Ants (mean \pm SD): *Tapinoma* sp. (Ti: $0.41 \pm 0.01 \text{ mm}, n = 30$), *Camponotus*

439 *nipponicus* (Cni: 0.77 ± 0.03 mm, n = 30), *Camponotus nawai* (Cna: 0.99 ± 0.03 mm, n = 100)

- 440 and *Camponotus devestivus* (Cd: 1.51 ± 0.10 mm, n = 3).
- 441

Thickness group	Diameter (mm)	Volume (103 cm^3)	Dry weight (kg)	D/V
Α	5.4 ± 2.3 (202)	2.1	1.41	0.67
В	13.0 ± 2.4 (159)	8.3	5.83	0.7
С	22.0 ± 3.1 (97)	13.9	8.87	0.64
D	28.4 ± 2.9 (66)	15.2	11.23	0.74
E	35.9 ± 3.3 (48)	17	12.18	0.71
F	48.3 ± 4.5 (32)	21.2	14.44	0.68
G	57.7 ± 3.2 (21)	19.9	12.48	0.62
н	78.4 ± 8.6 (12)	23.5	11.75	0.5

Table 1. Diameter and weight of each thickness group (A–H) of dead wood of ubame oak (*Quercus phillyraeoides*)

445 Mean diameter ± SD (number of pieces of dead wood), volume, dry weight and dry weight per

446 volume (D/V) are presented.

447

448 Table 2. Number of individuals of arthropods fauna in each thickness group (A–H) of dead wood of

449 ubame oak (*Quercus phillyraeoides*)

Thickness group	Α	в	С	D	E	F	G	н	Total
Insecta									
Orthoptera	-	-	-	1	-	-	-	-	1
Psocoptera	5	16	29	25	6	17	5	1	104
Thysanoptera	-	2	-	-	2	-	-	-	4
Hemiptera	-	-	-	-	-	_	3	-	3
Lepidoptera	-	-	-	2	12	25	48	144	231
Coleoptera	28	127	103	120	178	70	49	52	727
Diptera	-	-	-	-	-	1	-	1	2
Hymenoptera (wasps)	5	3	55	222	238	175	78	55	831
Hymenoptera (ants)	513	2772	2576	2598	828	1239	284	29	10839
Other arthropods									
Geophilomorpha	-	-	4	-	-	-	-	-	4
Polyxenida	18	112	188	86	17	8	1	1	431
Pseudoscorpiones	9	2	-	-	2	1	-	-	14
Acarina	2	4	8	-	-	-	-	1	15
Araneida	1	3	3	4	4	3	1	_	19
Collembola	-	11	35	2	-	_	-	-	48
Total	581	3052	3001	3060	1287	1555	469	284	13289

 $\begin{array}{c} 450\\ 451 \end{array}$

Thickness group	Family	Genus	Species	Individuals	Density (10^3 cm^-3)
А	3	4	4	28	13.3
В	3	7	7	127	15.3
С	5	9	9	103	7.4
D	5	11	11	120	7.9
E	6	11	11	178	10.5
F	7	11	11	70	3.3
G	9	13	13	49	2.5
н	9	13	14	52	2.2
Total	16	28	29	727	5.9

Table 3. Number of families, genus, species, individuals and density of beetles emerged from each

thickness group (A–H) of dead wood of ubame oak (*Quercus phillyraeoides*)

Thickness group	Subfamily	Genus	Species	Individuals	Density (10^3 cm^−3)
Α	2	2	3	513	6.5
В	3	3	4	2772	45.5
С	3	3	4	2576	72
D	3	4	5	2598	108.5
E	2	3	5	828	49.1
F	3	5	6	1239	106.4
G	2	2	3	284	37.5
Н	3	3	4	29	6.1
Total	3	6	8	10 839	89.1

456 **Table 4.** Number of subfamilies, genus, species individuals and density of arboreal ants emerged

	457	from each thickness gro	up (A–H) of dead wo	ood of ubame oak (<i>Quercu</i>	us phillyraeoides)
--	-----	-------------------------	---------------------	----------------------------------	--------------------

458

I